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Shell shape, allometry, and heavy metals content of two oyster species in the southeastern Gulf of California

Forma de la concha, alometría y contenido de metales pesados de dos especies de ostión en el sureste del Golfo de California

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# ABSTRACT

**Background.** The diverse and growing anthropogenic activity on the world's coastlines puts pressure on the shape and growth of the shells of different bivalve mollusks. Goals. Determine the biometric indicators and relative growth of the shell of Saccostrea palmula and Crassostrea corteziensis in wild populations of the southeastern Gulf of California. Methods. Oyster samplings were performed seasonally (summer 2019 to spring 2020) in the Altata (AL), Macapule (ML), Navachiste (NL) and El Colorado (ECL) lagoons, Sinaloa, Mexico (n = 300 per species and site), to document possible differences between shell measurements (length, height and width), biometric indices (elongation, compaction and convexity), and allometry, with the concentrations of the heavy metals (HM) such as arsenic (As); cadmium (Cd); copper (Cu); iron (Fe); lead (Pb); and zinc (Zn) found in the soft tissue of both species. Results. Except for height, the other shell dimensions and body weight presented significant differences (p < 0.05) between the sampling sites for each oyster. The greatest shell elongation of S. palmula and C. corteziensis was found in ML and ECL, respectively (p < 0.05). At each sampling site, the morphometric associations for the two oyster species showed a linear and positive trend, with different negative allometry (b < 1). The concentrations of organic and inorganic matter as well as HM analyzed (except zinc) showed correlations with some biometric indicators and allometry in most of the lagoons in the two species. Conclusions. The HM were related to the functional indicators of the shell and allometry of both ostreids in the different lagoons.

Keywords: Bivalves, biometric indexes, morphometry, anthropogenic activities, environmental variables,

## RESUMEN

Antecedentes. La diversa y creciente actividad antropogénica en los litorales del mundo ejerce presión en la forma y crecimiento de las conchas de diferentes moluscos bivalvos. Objetivos. Determinar los indicadores biométricos y crecimiento relativo de la concha de Saccostrea palmula y Crassostrea corteziensis en poblaciones silvestres del sureste del Golfo de California. Métodos. Se realizaron muestreos de ostiones estacionalmente (verano 2019 a primavera 2020) en las lagunas Altata (AL), Macapule (ML), Navachiste (NL) y El Colorado (ECL), Sinaloa, México (n = 300 por especie y sitio), para documentar posibles diferencias entre las medidas de la concha (longitud, altura y anchura), índices biométricos (elongación, compactación y convexidad), y alometría, con las concentraciones de los metales pesados (HM) como el arsénico (As); cadmio (Cd); cobre (Cu); hierro (Fe); plomo (Pb); y zinc (Zn) encontrados en el tejido blando de las dos especies. Resultados. Excepto para la altura, las demás dimensiones de la concha y el peso corporal presentaron diferencias significativas (p < 0.05) entre los sitios de muestreo para cada especie de ostión. La mayor elongación de la concha de S. palmula y C. corteziensis se encontró en ML y ECL, respectivamente (p < 0.05). En cada sitio de muestreo, las asociaciones morfométricas para las dos especies de ostión mostraron tendencia lineal y positiva, con diferente alometría negativa (b < 1). Las concentraciones de materia orgánica e inorgánica, así como de los HM analizados (excepto el Zn), mostraron correlaciones con algunos indicadores biométricos y alometría en los dos bivalvos, en la mayoría de las lagunas. Conclusiones. Los HM estuvieron relacionados con los indicadores funcionales de la concha y la alometría de ambos ostreidos en las diferentes lagunas.

Palabras clave: Bivalvos, índices biométricos, morfometría, actividades antropogénicas, variables ambientales.

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### INTRODUCTION

The biological synthesis of minerals for the formation of shells in bivalve mollusks is a process directed, first, by the genetic information of the organisms (Clark *et al.*, 2020) that triggers phenotypic expression (color, shape, structures, etc.) particular to each species. However, external factors such as food and environmental conditions –including anthropogenic activities that take place in the vicinity of the habitat (Grizzle *et al.*, 2017; Stewart *et al.*, 2021)– cause the shape of the shells as well as their growth to undergo variations that can be detected, even in populations of the same species that are close to each other. Due to the bio-mineralization of the shells, it is possible to infer the geo-climatic conditions to which they were exposed in their lives (Jacob *et al.*, 2008), which is why they represent excellent archives of biological-environmental information.

Shell measurements, commonly reported in field studies on bivalves (height, length, and width) are evaluation tools that, in conjunction with other analytical techniques, offer practical and immediate evidence of possible variations in the growth and shell shape of these invertebrates (Caill-Milly *et al.*, 2014; Dar *et al.*, 2018). Additionally, the elongation, convexity, and compaction of the valves in these mollusks are functional indicators that modify their proportion as a reflection of the adaptation to the conditions of a certain locality (Modestin, 2017), which, finally, are related to factors such as waves and currents (Telesca *et al.*, 2018), food availability (Nakano *et al.*, 2010), and anthropogenic activity (Stewart *et al.*, 2021). On the other hand, morphometric relationships between shell dimensions and body weight in bivalves are used to interpret their development, physiological condition (Zhang *et al.*, 2023), and relative growth, among one or several populations of the same species (Karakulak *et al.*, 2006).

The mangrove oysters *C. corteziensis* (Hertlein, 1951) and *S. palmula* (Carpenter, 1857) are bivalves of economic importance and aquaculture potential (Cáceres-Martínez *et al.*, 2012), with geographic coincidence in the Gulf of California (Lodeiros *et al.*, 2020). Both ostreids cohabit attached to the roots of mangroves, rocks exposed to low tides, and shells found in coastal lagoons. However, these ecosystems are influenced by anthropogenic activities, among which mining, aquaculture, and agriculture stand out as sources to the entrance of heavy metals (HM) (Páez-Osuna *et al.*, 2017). The effect that HM have on the bio-mineralization of the shell of some bivalve species has been highlighted, causing its thinning and easy fracture or breakage (Dar *et al.*, 2018; Stewart *et al.*, 2021), in places where there is an increasing contribution of these elements derived from intense anthropogenic activity, as is happening in the southeast of the Gulf of California (Páez-Osuna *et al.*, 2017).

Although there is information related to growth, survival, condition index, and environmental conditions in which these oyster species live (Cabrera-Peña *et al.*, 2001; Castillo-Durán *et al.*, 2010), as well as genetic variability (Pérez-Enríquez *et al.*, 2008; Mazón-Suástegui *et al.*, 2016), pathology (Cáceres-Martínez *et al.*, 2012; Villanueva-Fonseca *et al.*, 2020), morphometric relationships (Góngora-Gómez *et al.*, 2018), biochemical composition (Hurtado *et al.*, 2012), gonadal development (Chávez-Villalba *et al.*, 2008), reproductive biology in the wild (Rodríguez-Jaramillo *et al.*, 2008; Alvarado-Ruiz, 2018) and cultivated (Góngora-Gómez *et al.*, 2020); but there are no reports that relate the level of HM accumulated in the soft-tissue of oysters, with the variation in the shape of their shell and allometry -jointly- for populations within the Gulf of California.

Therefore, the objective of this work was to determine the biometric indicators and relative growth of the shell of *S. palmula* and *C. corteziensis* in wild populations of four coastal lagoons in the southeastern Gulf of California, to document possible differences based on the level of HM found in the soft tissue of the two species. It is inferred that, in general, the results will vary –even in the same species– since they depend on the particular conditions (environmental and anthropogenic) of each sampling site.

#### MATERIALS AND METHODS

Seventy-five specimens of each oyster species (*S. palmula* and *C. corteziensis*) were collected from mangrove roots during low tide in September (summer 2019), December (fall 2019), March (winter 2020), and June (spring 2020) in the Altata (AL; 24°36'89.6" N; 107°52'14.2" W), Macapule (ML; 25°24'35.9" N; 108°41'37.1" W), Navachiste (NL; 25°30'31.0" N; 108°50'95.4" W), and El Colorado (ECL; 25°46'14.5" N; 109°24'44.2" W) lagoons, from Sinaloa, Mexico (n = 300 per species and site). These coastal lagoons are located within the Gulf of California (Fig. 1) and are characterized by maintaining a permanent connection with the gulf through one or two mouths. In addition, they are surrounded by four species of mangrove (irregularly distributed), human settlements, agricultural crops, and shrimp farms (Páez-Osuna & Osuna-Martínez, 2015).

In each sampling, the environmental variables were recorded: water temperature and dissolved oxygen (D0) with an oximeter (YSI, 55/12 FT, Oxymeter, Ohio, USA); to measure salinity, a precision refractometer (ATAGO, S/Mill refractometer) was used; while the pH was obtained with a potentiometer (Hanna, HI 8314 pHmeter, USA) (Góngora-Gómez *et al.*, 2020). The concentrations of organic matter (OM) and inorganic matter (IM) were determined with the gravimetric method described by APHA (1995). The concentration of chlorophyll *a* (CI–*a*) was obtained with the spectrophotometric technique described by Strickland & Parsons (1972) and the equations of Jeffrey & Humphrey (1975).

The oysters were transported in a cooler with seawater ( $\approx 4 \,^{\circ}$ C) to the laboratory, where they were cleaned with plastic brushes to remove excess sediment and detach adhered shells, epibiont fauna, and mangrove remains (Sepúlveda *et al.*, 2023). A digital Vernier ruler (0.00 mm, Mitutoyo, CD-8" CS) was used to measure the length (SL; maximum distance between the anterior and posterior margins), height (SH; maximum distance from the umbo to the ventral margin), and shell width (SW; maximum distance between the thickest parts of the valves). Additionally, the oysters were dried with absorbent paper before weighing them to obtain the total body weight (BW), using a precision scale (0.001 g, OHAUS, Scout Pro SP 2001) (Rodríguez-Quiroz *et al.*, 2016). The shell measurements of each oyster were used to obtain the biometric indices: elongation (SH/SL), roundness or compaction (SW/SL), and convexity (SW/SH) (Selin, 2007; Modestin, 2017).

To determine the allometry of *S. palmula* and *C. corteziensis*, the measurements of SL, SH, and SW were considered. Outliers were removed from all data sets (Durbin-Watson test) and residuals were analyzed for normal distribution using quantile–quantile plot (RStudio, R Core Team 2018). The morphometric relationships between the diffe-

rent dimensions of the shell (SL/SH, SW/SL, and SW/SH, n = 300 per a principal composition species and site), for each coastal lagoon, were obtained with the linear lationship between

species and site), for each coastal lagoon, were obtained with the linear equation Y = bX + a; where Y and X = shell dimensions (SL, SH, and SW, mm); where a = intercept and b = slope. In these associations with the same unit of measurement, when the exponent b = 1, the morphometric relationships indicate isometric growth.

The concentrations of HM (arsenic, As; cadmium, Cd; copper, Cu; iron, Fe; lead, Pb; zinc, Zn) in the soft tissue of the oysters *S. palmula* and *C. corteziensis* were analyzed using atomic absorption spectrophotometry and the accuracy of the analytical method was evaluated with the standard reference material DOLT– $5^{\circ}$  (National Research Council Canada) mentioned in Sepúlveda *et al.* (2023).

The mean, standard deviation, minimum, and maximum values of shell dimensions and BW were reported by oyster species and sampling site. All data sets passed the assumptions of normality (Kolmogorov-Smirnov) and homoscedasticity (Bartlett). An analysis of variance and a Tukey test was performed to detect and highlight statistical differences between the oyster shell dimensions and the biometric indexes of each site ( $\alpha = 0.05$ ). The goodness of fit of the data was analyzed with the Pearson correlation coefficient (r) (Sokal & Rohlf 1995). Finally,

a principal component analysis (PCA) was applied to determine the relationship between the oysters' biometric and allometry variables and environmental variables by species and lagoon, with the level of HM found in the soft tissue. The STATISTICA 7 program (StatSoft, Tulsa, OK, USA) was used.

## RESULTS

The environmental variables in the sampling sites presented a fluctuation pattern, according to the annual seasons. The lowest values of water temperature, DO, salinity, pH, OM, IM, and Cl–*a* were found at the sites ECL (20.9 °C, autumn), ML (4.2 mg L<sup>-1</sup>, spring), AL (25.0 ‰, winter), ML (7.3, summer), ML (6.4 mg L<sup>-1</sup>, autumn), AL (19.0 mg L<sup>-1</sup>, autumn) and AL (1.6 mg m<sup>-3</sup>, autumn), respectively; while the highest values in AL (32.7 °C, spring), NL (8.7 mg L<sup>-1</sup>, winter), ECL (41.0 ‰, spring), NL (8.0, spring), AL (16.2 mg L<sup>-1</sup>, autumn), ECL (41.1 mg L<sup>-1</sup>, summer) and NL (9.3 mg m<sup>-3</sup>, winter), respectively (Table 1). Only salinity (*F* = 3.94, *p* = 0.03) and pH (*F* = 4.18, *p* = 0.03) showed significant differences (*p* < 0.05) concerning the sampling sites, with intervals of 30.5 ± 5.6 ‰ (AL) to 37.5 ± 2.9 ‰ (ECL) and from 7.5 ± 0.2 (ML) to 7.9 ± 0.1 (NL), respectively.



Figure 1. Location of sampling sites –Altata (AL), Macapule (ML), Navachiste (NL), and El Colorado (ECL) coastal lagoons– in the state of Sinaloa, Mexico (SE Gulf of California).

Sampling site	Temperature	DO	Salinity	рH	OM	IM	CI-a
	(°C)	(mg L-')	(‰)	•	(mg L-')	(mg L-')	(mg m <sup>-3</sup> )
Altata lagoon							
Mean $\pm$ SD	27.1 ± 5.9	$5.8 \pm 0.8$	$30.5 \pm 5.6^{a}$	$7.6 \pm 0.1^{a}$	$11.3 \pm 4.1$	$26.4 \pm 7.5$	2.7 ± 1.1
Min–Max	21.3-32.7	4.8-6.7	25.0-38.0	7.5-7.7	6.7–16.2	19.0–36.8	1.6-3.7
Macapule lagoon							
Mean $\pm$ SD	$26.4 \pm 5.7$	$5.4 \pm 0.7$	$31.0 \pm 2.7^{a}$	$7.5 \pm 0.2^{a}$	9.1 ± 2.8	$35.4 \pm 5.4$	3.9 ± 1.2
Min–Max	21.1-31.5	4.2-5.9	29.0-35.0	7.3-7.8	6.4–12.9	27.7-39.6	2.6-5.3
Navachiste lagoon							
Mean $\pm$ SD	$26.9 \pm 5.6$	$6.5 \pm 1.4$	$35.8 \pm 1.5^{ab}$	$7.9 \pm 0.1^{b}$	$9.2 \pm 0.7$	$32.9 \pm 4.7$	4.8 ± 3.2
Min–Max	21.4-31.8	5.4-8.7	35.0-38.0	7.8-8.0	8.2-9.8	26.1-36.5	2.3-9.3
El Colorado lagoon							
Mean $\pm$ SD	$26.8 \pm 6.3$	$6.1 \pm 0.5$	$37.5 \pm 2.9^{b}$	$7.6 \pm 0.1^{a}$	9.3 ± 2.1	$34.6 \pm 6.0$	4.4 ± 1.7
Min–Max	20.9-32.6	5.4-6.7	34.0-41.0	7.5-7.7	7.9–12.4	29.0-41.1	2.5-6.4

Table 1. Environmental variables in the sampling sites. Taken from Sepúlveda et al. (2023).

D0 = dissolved oxygen, OM = organic matter, IM = inorganic matter, CI-a = chlorophyll a, SD = standard deviation, Min = minimum, Max = maximum. Columns with different superscript letters denote significant differences (p < 0.05) among sampling sites.

The range of shell dimensions of *S. palmula* and *C. corteziensis* that were considered in this analysis (n = 1200 per species) was respectively 55.0–88.1 and 61.0–108.9 mm for SH, 32.3–68.0 and 32.1–84.2 mm for SL and 12.4–40.0 and 21.7–49.7 mm for SW; while, for the BW, it was 13.9–65.0 and 30.1–126.9 g (Table 2). Except for SH, the other shell dimensions and BW presented significant differences (p < 0.05) between the sampling sites for each oyster species.

The biometric indexes of the oyster shells in the four coastal lagoons were significantly different (p < 0.05). The greatest elongation of *S. palmula* and *C. corteziensis* was found in ML and ECL, respectively, while compaction and convexity were greater for both oyster species in AL (Table 3).

At each sampling site, morphometric associations for the two oyster species showed a linear and positive trend. In *S. palmula*, the determination coefficient (R<sup>2</sup>) ranged from 0.05 (*b* SW/SH in NL) to 0.55 (*b* SL/SH in AL) (Fig. 2), while for *C. corteziensis*, the lowest value (R<sup>2</sup> = 0.08) was obtained in ECL for the *b* SW/SH ratio, and the highest (R<sup>2</sup> = 0.70) was recorded in AL for *b* SL/SH (Fig. 3). In all sampling sites the morphometric relationships showed negative allometry (*b* < 1).

Table 4 shows the annual average values of the concentration of HM (As, Cd, Cu, Fe, Pb, and Zn) in the soft tissue of the two oyster species (taken from Sepúlveda *et al.*, 2023). Except for Cd in *S. palmula* and As in *C. corteziensis*, the concentrations of the other HM showed significant differences (p < 0.05) in the sampling sites for each oyster species. The highest levels of Cu, Pb, and Zn in the tissue of *S. palmula* and *C. corteziensis* were found in ECL, while Fe concentrations in both oyster species were higher in ML.

Sampling site	SH	SL	SW	BW
S. palmula				
AL	$70.16 \pm 4.07$	51.7 ± 6.07°	$28.5 \pm 4.77^{\circ}$	$36.75 \pm 10.7^{\text{b}}$
ML	$73.20 \pm 3.65$	$47.57 \pm 5.25^{a}$	$23.27 \pm 4.00^{a}$	35.37 ± 10.7 <sup>b</sup>
NL	$72.42 \pm 4.72$	$48.30 \pm 6.02^{a}$	26.67 ± 3.12 <sup>b</sup>	$33.55 \pm 5.50^{a}$
ECL	$74.77 \pm 3.70$	$50.25 \pm 4.92^{\text{b}}$	$26.32 \pm 4.30^{\circ}$	$38.7 \pm 8.35^{\circ}$
C. corteziensis				
AL	75.11 ± 4.91	$59.95 \pm 7.62^{\circ}$	37.35 ± 4.72°	$56.57 \pm 12.95^{a}$
ML	$74.75 \pm 4.67$	$54.55 \pm 7.12^{a}$	33.65 ± 3.85 <sup>b</sup>	59.45 ± 12.82 <sup>b</sup>
NL	73.05 ± 4.51	$56.52 \pm 6.60^{\text{b}}$	$34.07 \pm 4.17^{b}$	$55.65 \pm 13.07^{a}$
ECL	73.01 ± 3.42	$53.85 \pm 6.22^{a}$	$32.6 \pm 3.35^{a}$	59.65 ± 12.37 <sup>b</sup>

Table 2. Shell dimensions (mm) and body weight (BW, g) of *S. palmula* y *C. corteziensis* in the sampling lagoons (Altata lagoon, AL; Macapule lagoon, ML; Navachiste lagoon, NL; El Colorado lagoon, (ECL); Sinaloa, Mexico.

SH = Shell height, SL = Shell length, SW = Shell width. Columns with different superscript letters denote significant differences (p < 0.05) among sampling sites.



Figure 2. Morphometric relationships (n = 300 oysters per lagoon) among the shell dimensions of *S. palmula* sampled in four lagoons (AL, Altata lagoon; ML, Macapule lagoon; NL, Navachiste lagoon; ECL, El Colorado lagoon) from the southeast Gulf of California.  $R^2 = coefficient of determination$ .

OM and IM concentrations showed a correlation with some biometric indicators and allometry of *S. palmula* in three of the lagoons. OM was associated with compaction and *b* SL-SH (r = 0.99, p = 0.001and r = -0.97, p = 0.02, respectively) in ECL, while it only exhibited correlation with *b* SW-SH (r = 0.95, p = 0.04) in NL. For its part, the IM level was negatively correlated with the elongation values (r = -0.98, p = 0.01), *b* SW-SL (r = -0.98, p = 0.01), and *b* SW-SH (r = -0.99, p =0.001) in ML. Water temperature and salinity showed, respectively, a correlation with compaction (r = 0.98, p = 0.01 and r = -0.96, p = 0.03) in NL and ECL. In the case of *C. corteziensis*, OM was associated with *b* SW-SH (r = -0.95, p = 0.04) and convexity (r = 0.99, p = 0.001) in AL and NL, respectively; MI with elongation (r = 0.96, p = 0.03) in AL; DO with convexity and compactness (r = -0.99, p = 0.001 for both) in NL and ECL; and temperature, with convexity (r = 0.98, p = 0.01) in ML.

On the other hand, some HM showed a correlation with the biometric indicators and the allometry of the oysters in the different lagoons. The Pb level in *S. palmula* showed correlation with SL (r = 0.97, p = 0.02) and SH (r = 0.96, p = 0.03) in ML and ECL, respectively, while Cu was associated with relative growth (*b* SL-SH, r = 0.96, p = 0.03) in Table 3. Biometric indexes (annual mean ± standard deviation) of oyster shells in the sampling lagoons.

Sampling site	Elongation	Compactness	Convexity	
S. palmula				
AL	$1.40 \pm 0.13^{a}$	$0.58 \pm 0.10^{\circ}$	$0.40\pm0.06^{\circ}$	
ML	$1.54 \pm 0.15^{d}$	$0.49\pm0.09^{\rm a}$	$0.32\pm0.07^{\rm a}$	
NL	1.51 ± 0.16°	$0.56 \pm 0.08^{\circ}$	$0.37\pm0.05^{\text{b}}$	
ECL	$1.45 \pm 0.13^{\text{b}}$	$0.52\pm0.06^{\text{b}}$	$0.36\pm0.05^{\text{b}}$	
C. corteziensis				
AL	$1.32 \pm 0.10^{a}$	$0.64 \pm 0.09^{b}$	$0.48\pm0.06^{\text{d}}$	
ML	$1.41 \pm 0.17^{b}$	$0.62\pm0.08^{\text{ab}}$	$0.45\pm0.05^{\text{b}}$	
NL	$1.33 \pm 0.16^{a}$	$0.61 \pm 0.09^{a}$	$0.46\pm0.06^{\circ}$	
ECL	1.46 ± 0.18°	$0.61 \pm 0.07^{a}$	$0.42\pm0.05^{\rm a}$	

AL = Altata lagoon, ML = Macapule lagoon, NL = Navachiste lagoon, ECL = El Colorado lagoon. Columns with different superscript letters denote significant differences (p < 0.05) among sampling sites.

ECL. As and Cd concentrations in *C. corteziensis* tissue were correlated (r = 0.96, p = 0.03 for both), respectively, with *b* SW-SL and *b* SL-SH, in AL and ML. On the other hand, convexity was negatively related to Pb in *S. palmula* from ML (r = -0.99, p = 0.01) and ECL (r = -0.97, p = 0.02), while compaction was associated with Fe (r = -0.96, p = 0.03) in *C. corteziensis* from AL.

Of the variances of all the variables analyzed (24) in the four sites, the eigenvalues of three components satisfactorily explain their correlations. The points obtained in the PCA -for the two oysters- in the four lagoons show different groupings of biometric indexes and allometry, in relation to the levels of HM in the soft tissue of S. palmula (Fig. 4) and C. corteziensis (Fig. 5). The dispersion of the points for S. palmula presented an interval of -0.09 to 0.44 in AL, -0.08 to 0.35 in ML, -0.00007 to 0.42 in NL and -0.003 to 0.46 in ECL. In the case of C. corteziensis, the intervals per lagoon were: -0.007 to 0.35, -0.002 to 0.36, -0.01 to 0.43, and -0.01 to 0.42 in AL, ML, NL, and ECL, respectively. All HM were shown, to a different extent, to be associated with the functional indicators of the shell and allometry. Cd accumulated in the soft tissue of oysters was found sequestered to convexity and compaction in S. palmula from AL and ML and to elongation of C. corteziensis from ECL. On the other hand, this metal showed grouping with b SW-SL and b SL-SH of S. palmula in AL and ECL, and with b SL-SH of C. corteziensis from ML. For their part, Cd, Cu, and Zn were sequestered in *b* SW-SL and *b* SL-SH of *S. palmula* and with elongation and compaction in *C. corteziensis*, the three elements in ECL. As was the only element related to *b* SW-SL and *b* SL-SH of *S. palmula* in ECL.

#### DISCUSSION

The biometric and morphological analysis of the shell of bivalve mollusks is part of the knowledge necessary to know about the interference of the environment in their development. Despite the significant differences (p < 0.05) obtained in salinity and pH between the sampling sites, the average values of all environmental variables –by annual season– in the four coastal lagoons were found within the ideal range considered for the growth of both oyster species (Chávez-Villalba, 2014), and are consistent with those reported by other studies in the area (Páez-Osuna & Osuna-Martínez, 2015; Góngora-Gómez *et al.*, 2016).

Estuaries and coastal lagoons are transition zones between the ocean and the continent, which are exposed to abrupt changes in water variables caused, mainly, by the effect of the rain-evaporation relation ship with the depth of the body of water, and by the quantity and quality of discharges from rivers, irrigation canals, and industrial drains (Costa *et al.*, 2018; Omarjee *et al.*, 2021; Elegbede *et al.*, 2023). The OM



Figure 3. Morphometric relationships (n = 300 oysters per lagoon) among the shell dimensions of *C. corteziensis* sampled in four lagoons (AL, Altata lagoon; ML, Macapule lagoon; NL, Navachiste lagoon; ECL, El Colorado lagoon) from the southeast Gulf of California.  $R^2 = coefficient of determination$ .

and IM concentrations showed the greatest significant association in the biometric indexes and allometry of the two oyster species in the lagoons. The parameters obtained from the studied lagoons may be due to the residence time of the water and the intensity of the tides (Takasu et al., 2020), concentration and quality of particles (Middelburg & Herman, 2007), use of molecules in the first trophic levels (Hope et al., 2020), and contributions of anthropogenic material (Canuel & Hardison, 2016), among others, specific to each of them. The OM and IM suspended in the water column represent all the microcomponents (phytoplankton, dissolved organic and inorganic particles, and even toxic ones -such as HM- among others) that are filtered, absorbed, assimilated, and/or accumulated by bivalves in the soft tissue and/or shell (Qiao et al., 2022). Although there is a coincidence in the type of climate (Csa climate, subtropical with dry summer, Chen & Chen, 2013) and the short distance between the sampled lagoons ( $\approx$  200 km), the dimensions of the shells (SL, SW) and the growth allometric of each oyster species did not show similarity; which, in part, could be explained by various factors, such as the quantity and quality of the particles generated by the different activities located on the periphery and close to each sampling site. For example, LA receives the discharge of 98,518 ha of intensive agriculture, in addition to urban waste from human settlements (≈ 1,059,617 inhabitants) (Frías-Espericueta et al., 2018); while ML and NL suffer the impact of effluents from agriculture (119,994 ha) and aquaculture (18,735 ha) activities, along with urban waste from approximately 295,353 inhabitants (Páez-Osuna & Osuna-Martínez, 2015). In the case of ECL, agricultural activity (196,549 ha), aquaculture (12,639 ha), municipal waste -generated by nearly 450,000 inhabitants- and fishing and livestock operations in the area,

contribute strongly to the levels of organic particles. and suspended inorganic substances (Sepúlveda *et al.*, 2023); among the latter, the HM.

The oysters were selected with similar SH (Table 2), however, SL, SW, and BW recorded significant differences (p < 0.05) for each species between the lagoons; which, in turn, caused different values of elongation, convexity, and compaction (*p* < 0.05). While *S. palmula* and C. corteziensis were more elongated in different lagoons (ML and ECL, respectively), both species showed the greatest compaction and convexity in AL. Since both ostreids coexist attached to the mangrove root, exposed to the same environmental factors and those derived from the change of tides (desiccation, waves, currents, etc.), such biometric differences could be mostly attributed to their state of sexual maturation (Chong et al., 2020), genetic aspects and availability of food particles (Ballesta-Artero et al., 2018), that is, CI-a and seston. Specifically, OM and IM were correlated with some dimensions and biometric indicators of oysters; which may be due to the different contributions of nutrients and anthropogenic compounds that vary according to the urban and industrial activities surrounding each lagoon. Mazzola & Sarà (2001) established that the growth of the mussel Mytillus galloprovincialis and the clam Tapes sp. is influenced by the OM represented by phytoplankton, organic waste from the mollusks themselves, and surplus diets for fish dissolved in the water that they filter as food. Cugier et al. (2022) used a 3D ecosystem model (hydrodynamics, primary production, and individual growth) in Bourgneuf Bay, France, to evaluate the growth of the oyster Crassostrea gigas, concluding that IM concentration prevents food uptake and, therefore, the development of bivalves.



Figure 4. The PCA of *S. palmula* in the Altata (AL), Macapule (ML), Navachiste (NL), and El Colorado (ECL) lagoon. *b* SW/SH = shell width/height relationship; *b* SL/SH = shell length/height relationship; *b* SW/SL = shell width/length relationship; BW = body weight; CI-a = chlorophyll a; COMP = shell compactness; CONV = shell convexity; D0 = dissolved oxygen; ELONG = shell elongation; IM = inorganic matter; OM organic matter; pH = pH units; SAL = salinity; SH = shell height; SL = shell length; SW = shell width; T °C = temperature; As = arsenic; Cd = cadmium, Cu = copper; Fe = iron; Pb = lead; Zn = zinc.



Figure 5. The PCA of *C. corteziensis* in the Altata (AL), Macapule (ML), Navachiste (NL), and El Colorado (ECL) lagoon. *b* SW/SH = shell width/height relationship; *b* SL/SH = shell length/height relationship; *b* SW/SL = shell width/length relationship; BW = body weight; CI-a = chlorophyll a; COMP = shell compactness; CONV = shell convexity; D0 = dissolved oxygen; ELONG = shell elongation; IM = inorganic matter; OM organic matter; pH = pH units; SAL = salinity; SH = shell height; SL = shell length; SW = shell width; T °C = temperature; As = arsenic; Cd = cadmium, Cu = copper; Fe = iron; Pb = lead; Zn = zinc.

Because HM are important components of IM, oyster development could respond to their accumulated concentration. For example, the Pb level in *S. palmula* was associated with SL and SH in two lagoons (ML and ECL), while the Pb concentration obtained from the soft tissue of *S. palmula* was associated with low values of compaction and convexity. The same happened with the level of Fe and compaction in *C. corteziensis* from AL. HM's effect on the dimensions and development of bivalve shells has been reported. For example, Stewart *et al.* (2021) concluded that the continuous contribution of Zn, Cu, and Pb, from the mining industry of the Isle of Man (North Irish Sea), affects the development of the shell of the queen clam *Pecten maximus*, causing its weakness and thinning;

therefore, such HM are considered a threat to the aquaculture industry on the islands of Great Britain. On the other hand, Beeby *et al.* (2002) demonstrated experimentally that an increase in the uptake of Pb included in the diet of the garden snail *Helix aspera* causes a reduction in the mass of its shell because less calcium (Ca) and magnesium (Mg) are deposited in it, due to the extra energy expenditure that this snail must make to excrete excess Pb. The above would explain the partial effect of HM on the shell dimensions of both oyster species in the four lagoons studied. However, to conclude with certainty about the previous point, it is advisable –in parallel with the analysis of HM levels in the soft tissue of oysters– to know their concentrations in the water.

Table 4. Heavy metals (annual mean concentrations, mg/kg, w.w.) in the soft tissue of *S. palmula* and *C. corteziensis* from four coastal lagoons in the southeast Gulf of California. Taken from Sepúlveda *et al.* (2023).

Site	As	Cd	Си	Fe	Pb	Zn
S. palmula						
AL	4.18±0.05 <sup>b</sup>	$1.74 \pm 0.05$	15.13±0.15 <sup>ab</sup>	$19.53 \pm 0.37^{a}$	$1.30 \pm 0.01^{a}$	72.81±0.49 <sup>ab</sup>
ML	$3.52 \pm 0.06^{ab}$	$1.37 \pm 0.07$	23.01±0.16 <sup>b</sup>	35.36±0.5 <sup>b</sup>	$1.13 \pm 0.02^{a}$	$75.34 \pm 0.4^{ab}$
NL	$3.28 \pm 0.04^{a}$	$1.89 \pm 0.03$	$8.26 \pm 0.16^{a}$	$21.01 \pm 0.53^{a}$	$1.47 \pm 0.02^{ab}$	$55.85 \pm 0.47^{a}$
ECL	$3.6 \pm 0.06^{ab}$	1.36±0.08	33.99±0.13°	$30.08 \pm 0.47^{ab}$	1.74±0.01 <sup>b</sup>	92.15±0.64 <sup>b</sup>
C. corteziensis						
AL	3.71±0.06	$0.8 \pm 0.04^{a}$	$10.51 \pm 0.06^{ab}$	38.12±0.32 <sup>a</sup>	1.41±0.03 <sup>b</sup>	60.53±0.32 <sup>b</sup>
ML	$3.66 \pm 0.06$	$1.05 \pm 0.01^{a}$	12.63±0.07 <sup>b</sup>	56.44±0.44 <sup>b</sup>	$0.97 \pm 0.03^{a}$	47.43±0.44 <sup>b</sup>
NL	3.48±0.07	1.78±0.03 <sup>b</sup>	$3.89 \pm 0.05^{a}$	33.11±0.37ª	$1.27 \pm 0.02^{ab}$	$30.95 \pm 0.32^{a}$
ECL	3.81±0.09	$0.9 \pm 0.02^{a}$	27.73±0.09°	$38.09 \pm 0.38^{a}$	1.79±0.02 <sup>c</sup>	81.53±0.51°

AL = Altata lagoon, ML = Macapule lagoon, NL = Navachiste lagoon, ECL = El Colorado lagoon. Columns with different superscript letters denote significant differences (p < 0.05) among sampling sites.

While all morphometric relationships showed a linear and positive trend, the regression equations presented a consistent pattern of negative allometric type (b < 1). The above is coincident with adult oysters (SH > 70 mm); which, in addition, go through stages of sexual maturation and reproduction in an annual cycle, as pointed out by Mena-Alcántar et al. (2017) and Alvarado-Ruíz (2018) for C. corteziensis (SH > 57.1 mm) and *S. palmula* (SH > 42 mm), respectively. In this work, factors such as environmental variables (EI-Sayed et al., 2011), food availability (Lee et al., 2018), and metabolic energy expenditure in the reproductive process (Mann et al., 2014), among others, altered the proportionality of their allometry, generating moderate values of b < 0.80 for the two oyster species. Despite this, different morphometric interactions better described the relative growth for each species (R<sup>2</sup> SW/SL = 0.56 for S. palmula in LEC;  $R^2$  SH/SL = 0.70 for C. corteziensis in AL) in different locations. The above could be explained by 1) the inter-specific phenotype of their shells, and 2) the intra-specific effect of the anthropogenic activity of each lagoon. For the first, it is documented that S. palmula has a semi-circular and cupped shape, while the SH predominates in C. corteziensis, making it more elongated (Lodeiros et al., 2020). Regarding their specific form, there is evidence that some HM derived from various industries -such as Fe and manganese (Mn)can cause, respectively, changes in both the coloration and the degree of thinning in the shell of various mollusks (Krupnova et al., 2017). The above would partially explain the differences in the biometry and isometry of the shells in both oysters at the sampling sites.

Some elements -- such as Cu, Fe, and Zn-- are essential for mollusks at levels that do not exceed their metabolic demand (Singh et al., 2011). Concentrations higher than their physiological requirements are regulated through excretion (Regoli et al., 1991) or even incorporated into their shell (Dar et al., 2018). The results obtained associate some of the HM with allometric growth and biometric indicators of the shell of the two oyster species in different lagoons (Cu, in S. palmula from AL with b SL-SH; As and Cd, in C. corteziensis with b SW-SL and b SL-SH in AL and ML, respectively; Cd, in S. palmula from AL and ML, respectively with convexity and COMP; Cd, Cu, and Zn with elongation for C. corteziensis from ECL; As with S. palmula in ECL), without showing any specific trend in relation to the species or place. The above places the environmental conditions of each lagoon -including their anthropogenic activities- as a possible main cause of the differences in the shell shape and allometry of each ostreid. However, it is not possible to satisfactorily conclude that the load of these HM in the soft tissue of both species, has had any direct effect on their allometry and biometric indexes. It is recommended to analyze the HM not only in the water and sediment but also in the shell of the oysters to know their traceability in the different eco-biological compartments.

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#### REFERENCES

ALVARADO-RUIZ, C. 2018. Factores ambientales y madurez sexual de un banco de ostras *Saccostrea palmula* (Carpenter, 1857) Mollusca,

Bivalvia en bahía Culebra, Costa Rica. *Revista Cubana de Investiga*ciones Pesqueras 35 (1): 1-8.

- APHA (AMERICAN PUBLIC HEALTH ASSOCIATION). 1995. Standard methods for the examination of water and wastewater. American Public Health Association. Washington, D.C.
- BALLESTA-ARTERO, I., R. JANSSEN, J. VAN DER MEER & R. WITBAARD. 2018. Interactive effects of temperature and food availability on the growth of Arctica islandica (Bivalvia) juveniles. Marine Environmental Research 133: 67-77. DOI: 10.1016/j.marenvres.2017.12.004
- BEEBY, A., L. RICHMOND & F. HERPÉ. 2002. Lead reduces shell mass in juvenile garden snails (*Helix aspersa*). *Environmental Pollution* 120: 238-288. DOI: 10.1016/S0269-7491(02)00151-3
- CABRERA-PEÑA, J. H., M. PROTTI-QUESADA, M. URRIOLA-HERNÁNDEZ & O. SÁE-NZ-VARGAS. 2001. Crecimiento y madurez sexual de una población de *Saccostrea palmula* (Mollusca: Bivalvia), Costa Rica. *Revista de Biología Tropical* 49 (3–4): 877-882.
- CÁCERES-MARTÍNEZ, J., M. GARCÍA-ORTEGA, R. VÁZOUEZ-YEOMANS, T. PINEDA-GARCÍA, N. A. STOKES & R. B. CARNEGIE. 2012. Natural and cultured populations of the mangrove oyster *Saccostrea palmula* from Sinaloa, Mexico, infected by *Perkinsus marinus*. *Journal of Invertebrate Pathology* 110 (3): 215-325. DOI: 10.1016/j.jip.2012.03.019
- CAILL-MILLY, N., N. BRU, M. BARRANGER, L. GALLON & F. DAMICO. 2014. Morphological trends of four Manila clam populations (*Venerupis philippinarum*) on the French Atlantic coast: identified spatial patterns and their relationship to environmental variability. *Journal of Shellfish Research* 33 (2): 355-372. DOI: 10.2983/035.033.0205
- CANUEL, E. A. & A. K. HARDISON. 2016. Sources, ages, and alteration of organic matter in estuaries. *Annual Review of Marine Science* 8: 409-34. DOI: 10.1146/annurev-marine-122414-034058
- CASTILLO-DURÁN, A. J., J. CHÁVEZ-VILLALBA, A. ARREOLA-LIZÁRRAGA & R. BARRA-ZA-GUARDADO. 2010. Comparative growth, condition and survival of juveniles oysters *Crassostrea gigas* and *C. corteziensis* cultivated in summer and winter. *Ciencias Marinas* 36: 29-39.
- CHÁVEZ-VILLALBA, J., A. HERNÁNDEZ-IBARRA, M. R. LÓPEZ-TAPIA & J. MAZÓN-SUÁS-TEGUI. 2008. Prospective culture of the Cortez oyster *Crassostrea corteziensis* from northwestern Mexico: growth, gametogenic activity, and condition index. *Journal of Shellfish Research* 27: 711-720. DOI: 10.2983/0730-8000(2008)27[711:PCOTCO]2.0.C0;2
- CHÁVEZ-VILLALBA, J. 2014. Cultivo de ostión Crassostrea gigas: análisis de 40 años de actividades en México. Hidrobiológica 24: 175-190.
- CHEN, D. & H. W. CHEN. 2013. Using the Köppen classification to quantify climate variation and change: An example for 1901-2010. *Environmental Development* 6: 69-79. DOI: 10.1016/j.envdev.2013.03.007
- CHONG, J., Q. LI, C. XU & S. LIU. 2020. Seasonal variation in growth and gonadal development of the Pacific oyster "Haida No. 3" in relation to the nutritive components. *Journal of Fisheries of China* 44 (3): 411-418. DOI: 10.11964/jfc.20190111634
- CLARK, M. S., L. S. PECK, J. ARIVALAGAN, T. BACKELJAU, S. BERLAND, J. C. R. CAR-DOSO, C. CAURCEL, G. CHAPELLE, M. DE NOIA, S. DUPONT, K. GHARBI, J. I. HOFF-MAN, K. S. LAST, A. MARIE, F. MELZNER, K. MICHALEK, J. MORRIS, J. M. POWER, K. RAMESH, T. SANDERS, K. SILLANPÄÄ, V. A. SLEIGHT, P. J. STEWART-SINCLAIR,

K. SUNDELL, L. TELESCA, D. L. J. VENDRAMI, A. VENTURA, T. A. WILDING, T. YARRA & E. M. HARPER. 2020. Deciphering mollusc shell production: the roles of genetic mechanisms through to ecology, aquaculture and biomimetics. *Biological Reviews* 95: 1812-1837. DOI: 10.1111/ brv.12640

- COSTA, C. R., M. F. COSTA, D. V. DANTAS & M. BARLETTA. 2018. Interannual and seasonal variations in estuarine water quality. *Frontiers of Marine Sciences* 5: 301. DOI: 10.3389/fmars.2018.00301
- CUGIER, P., Y. THOMAS & C. BACHER. 2022. Ecosystem modelling to assess the impact of rearing density, environment variability and mortality on oyster production. *Aquaculture Environment Interactions* 14: 57-70. DOI: 10.3354/aei00428
- DAR, M. A., A. A. BELAL & A. G. MADKOUR. 2018. The differential abilities of some molluscs to accumulate heavy metals within their shells in the Timsah and the Great Nitter lakes, Suez Canal. *Egyptian Journal* of Aquatic Research 44: 291-298. DOI: 10.1016/j.ejar.2018.11.008
- ELEGBEDE, I., A. LAWALARE, O. FAVOUR, T. JOLAOSHO & A. GOUSSANOU. 2023. Chemical compositions of bivalves shells *Anadara senilis*, *Crassostrea gasar* and *Mytilus edulis* and their potential for a sustainable circular economy. *SN Applied Sciences* 5: 44. DOI: 10.1007/s42452-022-05267-7
- EL-SAYED, A. E. H., F. A. ABDEL-RAZEK, M. M. ABOU-ZAID & S. M. TAHA. 2011. Measures of allometric growth of black-lip pearl *Pinctada margaritifera* (Linnaeus 1758) Red Sea, Egypt. *International Journal of Zoological Research* 7 (2): 201-211. DOI: 10.3923/ijzr.2011.201.211
- FRÍAS-ESPERICUETA, M. G., A. VARGAS-JIMÉNEZ, J. RUELAS-INZUNZA, I. OSUNA-LÓPEZ, M. AGUILAR-JUÁREZ, J. C. BAUTISTA-COVARRUBIAS & D. VOLTOLINA. 2018. Total mercury in the mangrove oyster *Crassostrea corteziensis* of the subtropical Urías lagoon (NW Mexico). *Turkish Journal of Fisheries and Aquatic Sciences* 18: 853-858. DOI: 10.4194/1303-2712v18\_7\_03
- GÓNGORA-GÓMEZ, A. M., M. GARCÍA-ULLOA, M. ARELLANO-MARTÍNEZ, S. ABAD, A. DOMÍNGUEZ & J. PONCE-PALAFOX. 2016. Annual reproductive cycle and growth of the pen shell *Atrina maura* (Pteroidea: Pinnidae) on sand-bottom culture in the Ensenada Pabellones lagoon system, Gulf of California, Mexico. *Invertebrate Reproduction & Development* 60 (1): 28-38. DOI: 10.1080/07924259.2015.1126535
- GÓNGORA-GÓMEZ, A. M., A. L. DOMÍNGUEZ-OROZCO, B. P. VILLANUEVA-FONSECA, N. P. MUÑOZ-SEVILLA & M. GARCÍA-ULLOA. 2018. Seasonal levels of heavy metals in soft tissue and muscle of the pen shell Atrina maura (Sowerby, 1835) (Bivalvia: Pinnidae) from a farm in the Southeast coast of the Gulf of California Mexico. Revista Internacional de Contaminación Ambiental 34 (1): 57-68. DOI: 10.20937/RICA.2018.34.01.05
- GÓNGORA-GÓMEZ, A. M., C. H. SEPÚLVEDA, H. A. VERDUGO-ESCOBAR, O. ASTOR-GA-CASTRO, H. RODRÍGUEZ-GONZÁLEZ, A. L. DOMÍNGUEZ-OROZCO, J. A. HERNÁN-DEZ-SEPÚLVEDA & M. GARCÍA-ULLOA. 2020. Gonadal maturity of *Crassostrea corteziensis* cultivated in the Gulf of California. *Latin American Journal of Aquatic Research* 483: 381-395. DOI: 10.3856/vol48-issue3-fulltext-2422
- GRIZZLE, R. E., K. M. WARD, C. R. PETER, M. CANTWELL, D. KATZ & J. SULLI-VAN. 2017. Growth, morphometrics and nutrient content of farmed Eastern oysters, *Crassostrea virginica* (Gmelin), in New Hampshi-

re, USA. Aquaculture Research 48 (4): 1525-1537. DOI: 10.1111/ are.12988

- HOPE, J. A., J. HEWITT, C. A. PILDITCH, C. SAVAGE & S. F. THRUSH. 2020. Effect of nutrient enrichment and turbidity on interactions between microphytobenthos and a key bivalve: Implications for higher trophic levels. *Frontiers of Marine Science* 7: 695. DOI: 10.3389/ fmars.2020.00695
- HURTADO, M. A., I. S. RACOTTA, F. ARCOS, E. MORALES-BOJÓRQUEZ, J. MOAL, P. SOUDANT & E. PALACIOS. 2012. Seasonal variations of biochemical, pigment, fatty acid, and sterol compositions in female *Crassostrea corteziensis* oysters in relation to the reproductive cycle. *Comparative Biochemistry and Physiology Part B: Biochemistry and Molecular Biology* 163: 172-183. DOI: 10.1016/j.cbpb.2012.05.011
- JACOB, D. E., A. L. SOLDATI, R. WIRTH, J. HUTH, U. WEHRMEISTER & W. HOFMEISTER. 2008. Nanostructure, composition and mechanisms of bivalve shell growth. *Geochemica et Cosmochimica Acta* 72: 5401-5415. DOI: 10.1016/j.gca.2008.08.019
- JEFFREY, S. W. & G. F. HUMPHREY. 1975. New spectrophotometric equation for determining chlorophyll *a*, *b*, *c*1 and *c*2. *Plant Physiology and Biochemistry* 167: 194-204.
- KARAKULAK, F. S., H. ERK & B. BILGIN. 2006. Length-weight relationships for 47 coastal fish species from the northern Aegean Sea, Turkey. *Journal of Applied Ichthyology* 22: 274-278. DOI: 10.1111/j.1439-0426.2006.00736.x
- KRUPNOVA, T. G., I. V. MASHKOVAA, A. M. KOSTRYUKOVA, E. E. SCHELKANOVA & S. V. GAVRILKINA. 2017. Gastropods as potential biomonitors of contamination caused by heavy metals in South Ural lakes, Russia. *Ecological Indicators* 95: 1001-1007. DOI: 10.1016/j.ecolind.2017.12.005
- LEE, Y. J., E. HAN, M. J. WILBERG, W. C. LEE, K. S. CHOI & C. K. KANG. 2018. Physiological processes and gross energy budget of the submerged longline-cultured Pacific oyster *Crassostrea gigas* in a temperate bay of Korea. *PLoS ONE* 13 (7): e0199752. DOI: 10.1371/journal. pone.0199752
- LODEIROS, C., P. VALENTICH-SCOTT, J. CHÁVEZ-VILLALBA, J. M. MAZÓN-SUÁSTEGUI & J. M. GRIJALVA-CHON. 2020. Tropical and subtropical ostreidae of the American Pacifc: Taxonomy, biology, ecology, and genetics. *Journal* of Shellfsh Research 39 (2): 181-206. DOI: 10.2983/035. 039.0202
- MANN, R., M. SOUTHWORTH, R. B. CARNEGIE & R. K. CROCKETT. 2014. Temporal variation in fecundity and spawning in the eastern oyster, *Crassostrea virginica*, in the Piankatank river, Virginia. *Journal of Shellfish Research* 33 (1): 167-176. DOI: 10.2983/035.033.0116
- Mazón-Suástegui, J. M., N. TRABAL-FERNANDEZ, I. LEYVA-VALENCIA, P. CRUZ-HER-NANDEZ & H. LATISNERE-BARRAGAN. 2016. 28S rDNA as an alternative marker for commercially important oyster identification. *Food Control* 66: 205-214. DOI: 10.1016/j.foodcont.2016.02.006
- MAZZOLA, A. & G. SARÀ. 2001. The effect of fish farming organic waste on food availability for bivalve molluscs (Gaeta Gulf, Central Tyrrhenian, MED): stable carbon isotopic analysis. *Aquaculture* 192: 361-379. DOI: 10.1016/S0044-8486(00)00463-4
- MENA-ALCÁNTAR, M., O. I. ZAVALA-LEAL, C. A. ROMERO-BAÑUELOS, J. M. J. RUIZ-VE-LAZCO, J. T. NIETO-NAVARRO, D. PALACIOS-SALGADO & J. M. PACHECO-VEGA. 2017. Reproduction of Cortez oyster, *Crassostrea corteziensis*

(Hertlein, 1951) in a growing area in the central Mexican Pacific coast. *Latin American Journal of Aquatic Research* 45 (2): 485-490. DOI: 10.3856/vol45-issue2-fulltext-23

- MIDDELBURG, J. J. & P. M. J. HERMAN. 2007. Organic matter processing in tidal estuaries. *Marine Chemistry* 106 (1-2): 127-147. DOI: 10.1016/j.marchem.2006.02.007
- MODESTIN, E. 2017. Morphological variations of the shell of the bivalve *Lucina pectinata* (Gmelin, 1791). *Journal of Advance Biology* 10 (2): 2092-2107. DOI: 10.24297/jab.v10i2.6355
- NAKANO, D., T. KOBAYASHI & I. SACAGUCHI. 2010. Predation and depth effects on abundance and size distribution of an invasive bivalve, the golden mussel *Limnoperna fortunei*, in a dam reservoir. *Limnology* 11: 259-266. DOI: 10.1007/s10201-010-0314-4
- OMARJEE, A., S. TALJAARD, C. L. RAMJUKADH & L. VAN NIERKERK. 2021. pH variability in catchment flows to estuaries – A South African perspective. *Estuarine, Coastal and Shelf Science* 262: 107605. DOI: 10.1016/j. ecss.2021.107605
- PAEZ-OSUNA, F. & C. C. OSUNA-MARTÍNEZ. 2015. Bioavailability of cadmium, copper, mercury, lead, and zinc in subtropical coastal lagoons from the Southeast Gulf of California using mangrove oysters (*Crassostrea corteziensis* and *Crassostrea palmula*). Archives of Environmental Contamination and Toxicology 68: 305-316. DOI: 10.1007/ s00244-014-0118-3
- PÁEZ-OSUNA, F., S. ÁLVAREZ-BORREGO, A. C. RUIZ-FERNÁNDEZ, J. GARCÍA-HER-NÁNDEZ, M. E. JARA-MARINI, M. E. BERGÉS-TIZNADO, A. PIÑÓN-GIMATE, R. ALONSO-RODRÍGUEZ, M. F. SOTO-JIMÉNEZ, M. G. FRÍAS-ESPERICUETA, J. R. RUE-LAS-INZUNZA, C. R. GREEN-RUIZ, C. C. OSUNA-MARTÍNEZ & J. A. SÁNCHEZ-CA-BEZA. 2017. Environmental status of the Gulf of California: a pollution review. *Earth-Science Reviews* 166: 181-205. DOI: 10.1016/j.earscirev.2017.01.014
- PÉREZ-ENRÍQUEZ, R., S. ÁVILA & A. M. IBARRA. 2008. Population genetics of the oyster *Crassostrea corteziensis* (Hertlein) juveniles. *Aquaculture* 263: 199-210. DOI: 10.7773/cm.v34i4.1290
- QIAO, L., Z. CHANG, J. LI & T. LI. 2022. Selective feeding of three bivalve species on the phytoplankton community in a marine pond revealed by highthroughput sequencing. *Scientific Reports* 12: 6163. DOI: 10.1038/s41598-022-08832-7
- REGOLI, F., E. ORLANDO, M. MAURI, M. NIGRO & G. A. COGNETTI. 1991. Heavy metal accumulation and cSHium content in the bivalve *Donacilla cornea. Marine Ecology Progress Series* 74: 219-224.
- RODRÍGUEZ-JARAMILLO, C., M. A. HURTADO, E. ROMERO-VIVAS, M. RAMÍREZ, J. L. MANZANO & E. PALACIOS. 2008. Gonadal development and histochemistry of the tropical oyster, *Crassostrea corteziensis* (Hertlein, 1951) during an annual reproductive cycle. *Journal of Shellfish Research* 27 (5): 1129-1141. DOI: 10.2983/0730-8000-27.5.1129

- Rodríguez-Quiroz, G., M. García-Ulloa, A. L. Domínguez-Orozco, T. N. Valenzuela-Hernández, E. Nava-Pérez & A. M. Góngora-Gómez. 2016. Relación del crecimiento, condición y supervivencia del ostión del Pacífico *Crassostrea gigas* y las variables ambientales, cultivado en suspensión en el sistema lagunar Navachiste-Macapule, Sinaloa, México. *Revista de Biología Marina y Oceanografía* 51 (3): 541-551. DOI: 10.4067/S0718-19572016000300006
- SELIN, N. I. 2007. Shell form, growth and life span of *Astarte artica* and *A. borealis* (Mollusca: Bivalvia) from the subtidal zone of northeastern Sakhalin. *Russian Journal of Marine Biology* 33 (4): 232-237.
- SEPÚLVEDA, C. H., M. I. SOTELO-GONZALEZ, C. C. OSUNA-MARTÍNEZ, M. G. FRÍAS-ES-PERICUETA, R. SÁNCHEZ-CÁRDENAS, M. E. BERGÉS-TIZNADO, A. M. GÓNGO-RA-GÓMEZ & M. GARCÍA-ULLOA. 2023. Biomonitoring of potentially toxic elements through oysters (*Saccostrea palmula* and *Crassostrea corteziensis*) from coastal lagoons of Southeast Gulf of California, Mexico: health risk assessment. *Environmental and Geochemistry Health* 45: 2329-2348. DOI: 10.1007/s10653-022-01347-0
- SINGH, R., N. GAUTAM, A. MISHRA & R. GUPTA. 2011. Heavy metals and living systems: An overview. *Indian Journal of Pharmacology* 43 (3): 246-253. DOI: 10.4103/ 0253-7613.81505
- SOKAL, R. R. & F. J. ROHLF. 1995. Biometry. W. H. Freedman and Company. New York.
- STEWART, B. D., S. R. JENKINS, C. BOIG, C. SINFIELD, K. KENNINGTON, A. R. BRAND, W. LART & R. KRÖGER. 2021. Metal pollution as a potential threat to shell strength and survival in marine bivalves. *Science of Total Environment* 755: 143019. DOI: 10.1016/j.scitotenv.2020.143019
- STRICKLAND, J. D. & T. R. PARSONS. 1972. A practical handbook for the seawater analysis. Bulletin of Fisheries Research Board of Canada.
- TAKASU, H., K. UCHINO & K. MORI. 2020. Dissolved and particulate organic matter dynamics relative to sediment resuspension induced by the tidal cycle in minotidal estuaries, Kyushu Japan. *Water* 12 (9): 2561. DOI: 10.3390/w12092561
- TELESCA, L., K. MICHALEK, T. SANDERS, L. S. PECK, J. THYRRING & E. M. HARPER. 2018. Blue mussel shell shape plasticity and natural environments: a quantitative approach. *Science Reports* 8 (1): 1-15. DOI: 10.1038/ s41598-018-20122-9
- VILLANUEVA-FONSECA, L. C., M. GARCÍA-ULLOA, M. LÓPEZ-MEYER, B. P. VILLANUE-VA-FONSECA, J. A. HERNÁNDEZ-SEPÜLVEDA, N. P. MUÑOZ-SEVILLA & A. M. GÓN-GORA-GÓMEZ. 2020. *Perkinsus marinus* in the pleasure oyster *Crassostrea corteziensis* cultivated on the southeast coast of the Gulf of California, Mexico. *Latin American Journal of Aquatic Research* 48 (4): 529-537. DOI: 10.3856/vol48-issue4-fulltext-2463
- ZHANG, Z., Y. WU, Y. ZHANG, Y. ZHU, Y. CAO, S. CHEN, Y. PENG, X. SUN & A. CHEN. 2023. Correlation of morphometric properties to meat yield and fatness index in the red strain of the saltwater hard clam *Meretrix meretrix. PLoS ONE* 18 (4): e0284730. DOI: 10.1371/journal. pone.0284730